

Assessment of Heavy Metal Levels in Water and Their Toxicity in Some Tissues of Nile Tilapia (*Oreochromis niloticus*) in River Nile Basin at Greater Cairo, Egypt

¹Abeer M. Badr, ¹Noha A. Mahana and ²Alaa Eissa

¹Zoology Department, Faculty of Science, Cairo University, Giza, 12613, Egypt

²Department of Fish Diseases and Management,
Faculty of Veterinary Medicine, Cairo University, Giza, 11221, Egypt

Abstract: The levels of lead (Pb), copper (Cu), cadmium (Cd), iron (Fe), zinc (Zn) and manganese (Mn) were measured in water samples and in tissues (liver, kidney, gill, spleen and muscle) of Nile tilapia collected from two locations along the River Nile basin at Greater Cairo in Egypt. We investigated the impact of these traces of heavy metals on the physiology, immune responses, genotoxicity and histology of Nile tilapia. The concentrations of heavy metals in water samples from polluted area (El-Tebeen, Area 2) were significantly higher than reference area (El-Zamalek, Area 1) for all investigated metals. The prevalence of each element in various fish tissues was variable between the two sampling areas. The level of plasma total proteins was significantly ($P < 0.001$) elevated in Area 2 compared to Area 1. Lysozyme levels were insignificantly declined in plasma of fish from Area 2 while phagocytosis percentage was markedly reduced ($P < 0.001$) in Area 2 compared to Area 1. DNA damage (comet assay) in blood elevated statistically in tilapia collected from Area 2 compared to Area 1. Additionally, numerous histopathological alterations were observed in Area 2. Thus, innate immune responses of tilapia could be sensitive to environmental contamination.

Key words: Immune Response • Innate Immunity • DNA Damage • Total Protein • Lysozyme Activity • Phagocytosis • Water Pollution

INTRODUCTION

The River Nile is the principal freshwater resource for Egyptians and represents more than 97% of Egyptian water resources [1]. Globally, industrial waste water represents the main source of water pollution. The Uprising increase in modern industries, agriculture urbanization, tourism and human activities are the main sources for chemical pollution to both, aquatic environment and its coexisting ecosystems [2]. Heavy metals are persistent contaminants in the environment causing serious illness in fish, animals and human. Regionally, industrial and agricultural runoffs are considered the primary source of metal poisoning to fish and other aquatic animals in Egypt [3,4].

Fish, the first vertebrate in the evolution tree, exhibits both innate and acquired immune responses. The immune system of fishes on contrary to higher vertebrates is comparatively simple and undifferentiated.

The major lympho-myeloid organs of fishes are thymus, head kidney and spleen [5]. The innate immunity is a fundamental defense mechanism of fish [6]. The immune system of aquatic organisms, such as fish, is continuously affected by periodic or unexpected changes of their environment. Adverse environmental situations may acutely or chronically stress the health of fish, altering some of their biochemical parameters and suppressing their innate and adaptive immune responses [4,7].

Tilapia (*Oreochromis (O.) spp.*) is a teleost fish with a worldwide distribution; therefore it is a good model for assessing the impacts of different environmental pollutants on aquatic ecosystems. A comparative study of five economically important taxa of tilapia showed that Nile tilapia (*O. niloticus*) presents a Godgifted strong immune system that maximizes their capability to tolerate biotic and abiotic types of stress [8,9]. Further, the natural surface feeder omnivorous non predator behavior of Nile

tilapia might grant them another natural tool that minimizes their possibility of contracting numerous types of pollutants including biological/chemical forms when compared to bottom feeder fishes [4,9]. In fish, chemical pollutants are commonly leading to genotoxicity through their aquatic environments [10].

In rivers, fish are often at the top of the food chain and have the tendency to accumulate heavy metals from water [11]. Therefore, bioaccumulation of metals in fish can be considered as an indicator of metal pollution in aquatic environment [12]. Prolonged exposure to water pollutants even in very low concentrations have been reported to induce morphological, histological and biochemical alterations in fish tissues, which may critically influence the quality and marketability of fishes [3,13]. However, their persistence in the aquatic environment is surely related to the success of their immune system to counteract such impacts [14]. It is also possible that environmental toxicants may increase the susceptibility of aquatic animals to various diseases by interfering with the normal functioning of their immune, reproductive and developmental processes [4].

The present study was carried out to determine the levels of bioaccumulation of some heavy metals [lead (Pb), copper (Cu), cadmium (Cd), iron (Fe), zinc (Zn) and manganese (Mn)] in aquatic environment and in some tissues of Nile tilapia. Another critical goal is to determine the impacts of these heavy metals on the histological normogram of Nile tilapia collected from possibly polluted and non-polluted areas of River Nile basin at the vicinity of Greater Cairo. Ultimately, our main goal was to assess the effects of these heavy metals on the innate immune responses and DNA integrity of collected Nile tilapias.

MATERIALS AND METHODS

Study Areas: Samples were collected from diverse geographical locations along the River Nile basin at the vicinity of two cities of Greater Cairo.

Study Area 1: The River Nile basin is at El-Zamalek district, this area can be considered as a reference control area due to its distance from any source of industrial/agricultural runoffs). This area located downstream in main River Nile extension at Greater Cairo. It is 37 km south to the industrial zone where no industrial effluent was recognized and thus considered as control area. Geographically, the reference area (Area 1) is located at GPS reading of 30°22'20.24"N and 31°13'10.23"E.

Study Area 2: The River Nile basin is at El-Tebeen district, this area located upstream south of River Nile extension at Greater Cairo. This Area is located at core of Cairo industrial zone where cement factory, electric power station, Iron/ steel factory dump their effluents directly to the stream of River Nile. Geographically, Area 2 is located at Global Positional System (GPS) reading of 29°41'21.03"N and 31°16'57.23"E.

Water and Fish Sampling: Water samples were collected manually in triplicates from each of the studied areas between 10:00 and 12:00 a.m. at a depth of 30 cm below the water surface and stored at 4°C in sterile glass bottles, acidified with concentrated hydrochloric acid (HCl) for preservation. Fifty adult Nile tilapias (*O. niloticus*) ranged from 250±10 g weight were collected during summer season in 2013 from both sampling areas. Fish were collected in closed mesh before transferred to the laboratory in well aerated Styrofoam boxes containing Nile water. Blood samples were withdrawn from the caudal vessels of collected fishes using heparinized syringes [15]. Portion of fresh blood were immediately used for comet assay and phagocytosis while the remaining part was centrifuged to obtain plasma. Collected plasma was kept at -20°C for further analysis to estimate lysozyme, total proteins, albumin and total globulins. Fishes were further dissected using three line incision [15] in order to get spleens, livers, kidneys, gills and muscles. Each tissue sample was divided into two halves. One half was kept frozen at -20°C for assessing heavy metals concentrations and the second half was fixed in 10 % buffered neutral formalin for histological examination.

Heavy Metals Assessment: Trace elements were measured in water and tissue samples using flame atomic absorption spectrophotometer (Thermo Scientific, UK) with double beam and deuterium background corrector according to APHA [16]. Tissue samples were dried, acid-digested and diluted with de-ionized water to known volume using the dry-ashing procedure proposed by Hseu [17]. Analytical blanks were run in the same way as the samples and concentrations were determined using standard solutions prepared in the same acid matrix. All reagents used were of analytical grade (Merck, USA). Standards for instrument calibration were prepared on the basis of mono-element certified reference solution inductively coupled plasma standard (Merck). Standard reference material (National Institute of Standards and Technology [NIST], USA) was used to validate analysis and the metal recoveries ranged between 90 and 110%.

Concentrations of all heavy metals in water are expressed as mg/l (ppm) and $\mu\text{g/g}$ wet weight in tissues.

Total Protein Content: Total protein was estimated in plasma according to Burtis *et al.* [18]. Total globulins were calculated by subtracting albumin from total protein.

Phagocytic Activity: Polymorphonuclearleukocytic cells (PMN) were isolated from blood using the method described by Rouse *et al.* [19]. The mixture of PMN and bacteria (*Staphylococcus aureus*) were incubated at 37°C for 2 h with regular stirring and then the mixture was centrifuged at 20000 g for 5 min at 4°C. The supernatants were used to estimate the percentage of bacteria phagocytosed. The mixture of bacteria and PMN were treated with one cycle of freezing and thawing. The percentage of bacteria killed was estimated according to the formula described by Woldehiwet and Rowan [20].

Plasma Lysozyme Activity: Plasma lysozyme activity was measured using the method described by Schultz [21]. Lysozymes diffused through the agarose gel containing a suspension of *Micrococcus lysodeikticus* (Sigma-Aldrich). A clear zone ring of lysis developed in the translucent agarose gel. The wells were swiftly filled with 25 μl aliquots from plasma samples. The plates were then covered tightly and incubated at room temperature on a level surface for 12-18 h. At the end of the incubation period, the clear zone ring diameters were measured to the nearest 0.1 mm.

Comet Assay: DNA damage was determined by alkaline comet assay. Comet assay was performed according to the method described by Tice *et al.* [22]. Ten μl of whole blood was mixed with 75 μl of 0.5% low-melting agarose and spread on pre-coated slides with normal melting agarose (1%). Slides were immersed in cold lysis buffer (2.5 M NaCl, 100 mM EDTA and 10 mM Tris-HCl, pH 10, with freshly added 10% DMSO and 1% Triton X-100) for 24 h at 4°C in darkness. Subsequently, the slides were incubated in fresh alkaline buffer for 20 min (300 mM NaOH and 1 mM EDTA, pH>13) and then electrophoresed for 20 min at 25 V and 300 mA. Neutralization with excess amount of 0.4 M Trizma base (pH 7.5), fixation in 100% cold ethanol and air drying were done. Finally, slides were stained with ethidium bromide (2 $\mu\text{g/ml}$). The extent of DNA migration for each sample was determined by simultaneous image capture and scoring of 100 cells at 400x magnification using Komet 5 image analysis software developed by Kinetic Imaging,

Ltd. (Liverpool, UK). The measured parameters were tail length, the %tail DNA and tail moment.

Histological Examination: The Nile tilapia tissues (gills, kidney, liver, spleen and muscles) were preserved in 10% buffered formalin, embedded within paraffin, sectioned at 5 μm and stained with Hematoxylin and Eosin (H&E) according to the method described by Prophet *et al.* [23]. The histological alterations of H&E stained tissue sections were microscopically assessed under low / high microscopic powers.

Statistical Analysis: The present data were analyzed by Student's "t" Test to compare different parameters between the two studying areas using SPSS statistical package version 20. The data values were expressed as mean \pm standard error of mean (SEM). $P < 0.05$ was the accepted significance level. Correlation coefficient (r) was used to fit the relationship between the concentration of heavy metals in water and various tissues of collected fish.

RESULTS

Heavy Metals Assessment: Analysis of water samples collected from the two River Nile sites has indicated that the concentrations of heavy metals in water samples from Area 2 were significantly higher for all investigated metals than Area 1. Briefly, Fe showed the highest mean level in polluted Area (Area 2), followed by Pb, Zn, Mn, Cd and Cu while in the reference area (Area 1), the order was Fe followed by Zn then Pb then Mn then Cd and finally Cu (Table 1).

Results of the assessment of the heavy metal residues in different edible (musculature) and non-edible fish tissues (gill, liver, kidney and spleen) of Nile tilapias collected from both studied areas were tabulated in Table 2. The recorded data indicated the presence of heavy metals traces in all tested fish tissues. Also, the prevalence of each element was variable between the two sampling sites. Gills contaminated with traces of heavy metals with the following order: Fe > Zn > Mn > Cu > Pb > Cd, whereas kidneys Fe > Zn > Pb > Cu > Mn > Cd. Splenic tissue of fishes from polluted area showed marked higher concentrations of Cu and Fe compared to the reference site. However, liver tissues demonstrated no significant changes between the two sampling sites except for Fe. Interestingly, the levels of heavy metals residues in muscle tissues revealed significant higher amounts of Fe, Zn, Cu, Cd and Pb at the polluted area compared to the

Table 1: The concentration of Zn, Mn, Cu, Cd, Pb and Fe (mg/l) in water samples of Nile River surface water collected from areas 1 and 2.

Metals	Water samples	
	Area 1	Area 2
Zn	0.1780±0.0090	0.2680±0.0060*
Mn	0.0039±0.0003	0.0466±0.0010**
Cu	0.0009±0.0001	0.0043±0.0006*
Cd	0.0038±0.0004	0.0150±0.0002**
Pb	0.0081±0.0001	0.3548±0.0080**
Fe	0.2396±0.0026	3.0500±0.1305**

Data are represented as mean± standard error of mean (SEM).
*, **: $P < 0.01$ and $P < 0.001$: significant difference in comparison with the reference standard area at $\alpha = 0.01$ and 0.001 , respectively.

Table 2: The concentrations of heavy metals in livers, kidneys, spleens, gills and muscles ($\mu\text{g/g}$ wet weight) of *O. niloticus* collected from reference and polluted areas.

Metals	Areas		
	Reference area	Polluted area	%Change
Zn			
Liver	24.03±2.67	20.48±1.91	-14.77
kidney	91.64±1.58	39.58±3.21***	-56.80
Spleen	39.62±8.86	49.85±9.70	+25.82
Gills	8.87±1.54	18.51±0.33***	+108.68
Muscles	0.90±0.22	6.99±0.31***	+676.66
Mn			
Liver	1.38±0.55	1.25±0.49	-9.42
kidney	0.70±0.38	3±0.59**	+328.57
Spleen	0.70.38	3.51±1.20	+401.42
Gills	3.69±0.22	5.61±1.63*	+52.03
Muscles	0.232±0.067	0.151±0.044	-34.91
Cu			
Liver	27.43±8.66	24.03±2.67	-12.39
kidney	8.16±2.12	91.79±2.03***	+1024.87
Spleen	8.42±1.86	31.17±3.43***	+270.19
Gills	1.91±0.21	8.86±1.99**	+363.87
Muscles	0.264±0.023	0.693±0.091**	+162.5
Cd			
Liver	0.07±0.04	0.17±0.06	+142.85
kidney	0.081±0.025	1.07±0.29*	+32.09
Spleen	0.14±.07	1.83±0.85	+1207.14
Gills	0.02±0.003	0.09±0.03	+350
Muscles	0.009±0.002	0.024±0.005*	+2.56
Pb			
Liver	2.25±0.19	4.28±1.01	+90.22
kidney	20.21±7.81	46.52±9.69	+130.18
Spleen	24.37±3.77	59.49±14.61	+144.11
Gills	1.02±0.06	1.66±0.19*	+62.74
Muscles	0.662±0.058	0.833±0.057*	+25.83
Fe			
Liver	24.33±2.25	104.39±15.69**	+329.05
kidney	131.54±20.25	174.35±26.38	+4.18
Spleen	290.31±49.91	496.99±36.18**	+71.192
Gills	49.95±8.6	101.05±11.13**	+102.30
Muscles	5.37±0.5	31.53±4.48***	+487.15

Data are represented as Mean ± standard error of mean (SEM).
*, **, ***: significant difference in comparison with the reference area at $\alpha = 0.05, 0.01, 0.001$ and 0.0001 , respectively.

Table 3: Relationship between the concentrations of metals (Zn, Cu, Mn, Cd, Pb and Fe) in water (mg/l) and in some tissues ($\mu\text{g/g}$ wet weight) of *O. niloticus* collected from polluted area.

	Heavy metals in water (mg/l)					
	Zn	Cu	Mn	Cd	Pb	Fe
Liver	+0.902	+0.264	+0.995	-0.153	-0.772	+0.010
Kidney	+0.837	+0.999	+0.975	+0.028	-0.182	-0.699
Spleen	+0.995	+0.085	+0.993	+0.388	+0.998	+0.530
Gills	+0.747	+0.850	+0.863	+0.406	+0.939	-0.968
Muscles	+0.998	-0.941	+0.989	+0.912	+0.373	+0.368

Pearson correlation coefficient (r)

reference area with a decreasing order of $\text{Fe} > \text{Zn} > \text{Pb} > \text{Cu} > \text{Mn} > \text{Cd}$ at both sites. The splenic tissue revealed the highest percentage of change for Cd, Pb and Mn while muscle tissues recorded the highest percentage of change for Zn and Fe and kidney tissues for Cu.

Correlation Between Heave Metals Levels in Water and Tissue Samples:

The Pearson Coefficient has indicated that the concentrations of different traces of heavy metals in various tissues of fish caught from polluted area were greatly dependent on the concentrations of these elements in the raw water (Table 3). The relationship between heavy metals concentrations in fish organs and external water environment were variable between positive and negative correlations. As demonstrated in Table 3, in polluted site, levels of Zn and Mn in water exhibited positive correlation with those in all tissues of *O. niloticus*. In addition, Cu in water showed a strong positive correlation with Cu in kidney and gills, despite the strong negative correlation in muscles. Ultimately, concentrations of Cd in water exhibited a strong positive correlation with its concentrations in muscles, while Pb showed a strong positive correlation with that in spleen and gill tissue.

Total Proteins:

The level of plasma total proteins was significantly ($P < 0.001$) elevated in the polluted industrial area compared to the reference area as shown in Fig. 1. The albumin mean level was not significantly different between the two sites. Further, the mean level (3.29 ± 0.1) of total globulins was significantly reduced ($P < 0.001$) in polluted area compared to reference area (7.15 ± 0.06).

Plasma Lysozyme Activity:

Analysis has indicated that lysozyme levels were insignificantly declined in plasma of fishes from polluted area compared to that of reference area. The mean level (90.09 ± 3.66) of lysozyme in Area 1 was not significantly different from that of fish in Area 2 (96.522 ± 2.21) (Fig. 2A).

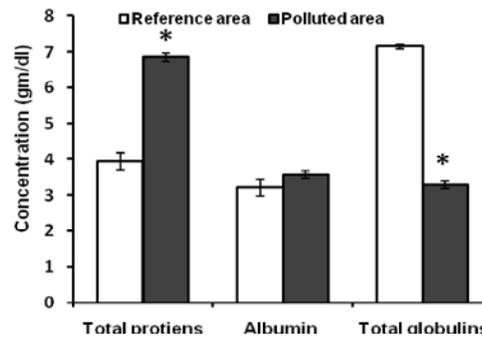


Fig. 1: Estimation of total proteins: Concentrations of total proteins, albumin and total globulins in plasma of *O. niloticus* collected from reference and polluted areas. *: significant difference in comparison to reference area at $\alpha = 0.0001$.

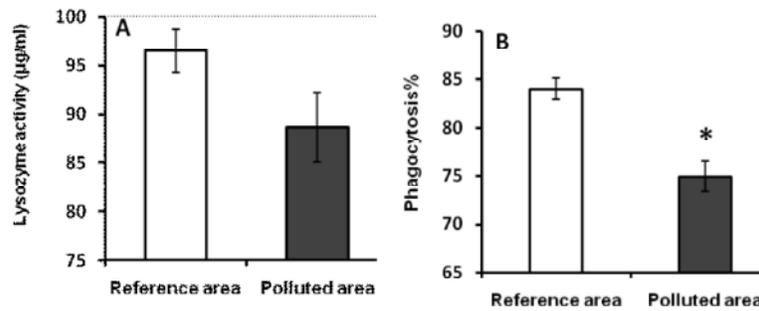


Fig. 2: Lysozyme activity and percentage of phagocytosis: (A) lysozyme activity in plasma of *O. niloticus* collected from reference and polluted areas. (B) percentage of phagocytosis of blood cells of *O. niloticus* collected from reference and polluted areas. *: significant difference compared to reference site at $\alpha = 0.0001$.

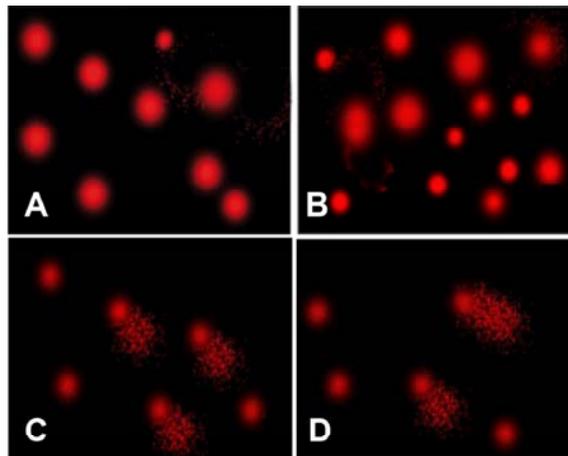


Fig. 3: Comet assay showing DNA distribution: (A & B) photomicrographs showed DNA damage of blood cells of Nile tilapia collected from reference area (Area 1). (C, D) Photomicrographs showed DNA damage of blood cells of Nile Tilapia collected from polluted area (Area 2).

Phagocytosis: Ingestion of bacteria is a method for evaluating the effect of heavy metals residues on phagocytes. The results of ingestion of bacteria by leukocytes showed a significant ($P < 0.001$) reduction (75.0 ± 1.236) in fish collected from polluted area compared to reference area (84.0 ± 1.64) (Fig. 2B).

Comet Assay: The Photomicrographs of DNA damage were presented in Fig. 3. The relative amounts of DNA strand breaks were higher in the polluted area (Area 2) than the reference area (Area 1) as shown in Fig. 4. The comet assay revealed a significant ($P < 0.01$) increase of DNA tail lengths (5.71 ± 0.35) in blood cells of *O. niloticus*

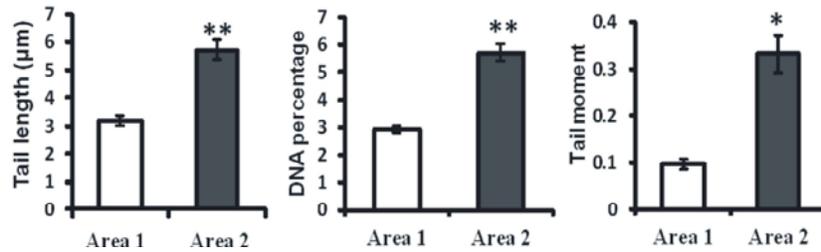


Fig. 4: Tail length, DNA percentage and Tail moment of comet parameters of blood cells of *O. niloticus* collected from reference (Area 1) and polluted (Area 2) areas. The data are represented as Mean ± Standard Error of Mean (SEM); *, **: significant difference in comparison to reference area at $\alpha=0.01$ and 0.0001 , respectively.

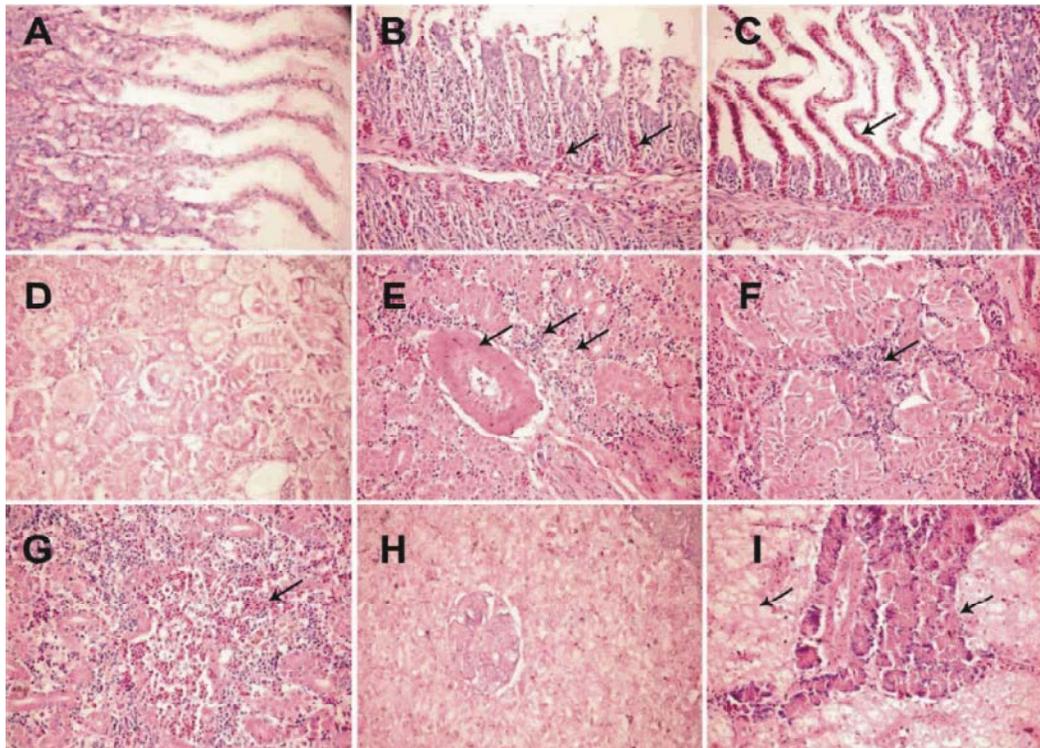


Fig. 5: Histological appearance of different tissues in *O. niloticus*. Photomicrographs of the gill tissue; (A) normal histological structure (B) necrosis (C) lamellar congestion. Photomicrographs of the kidney tissue; (D) no histological changes (E) showed marked interstitial nephritis and hyalinosis in the wall of renal blood vessels (F) focal tubular necrosis associated with leucocytic cells infiltration. (G) Photomicrograph of kidney showing massive inflammatory cells and free RBCs. Photomicrographs of liver tissue; (H) normal architecture (I) necrosis of hepatopancreas and vacuolization of hepatocytes. All tissues stained with hematoxylin and eosin (x400). A, D and H photomicrographs represented reference area, while other photomicrographs represented polluted area.

collected from polluted area compared to that of reference area (3.21 ± 0.17). The percentage of DNA also elevated significantly ($P < 0.01$) in polluted area (5.71 ± 0.32) compared to the reference area (2.95 ± 0.14). *O. niloticus* from polluted area in Nile River had blood cells with higher significant ($P < 0.0001$) DNA tail moments (0.332 ± 0.039) than referenced area (0.098 ± 0.011).

Histological Findings: The histological examination of different sampled organs of collected Nile tilapias has revealed the presence of numerous histopathological alterations among examined muscles, livers, gills and kidney tissues (Figs. 5, 6). The gills from the reference area had normal structure as shown in Fig. 5A, while gills from the polluted area have presented

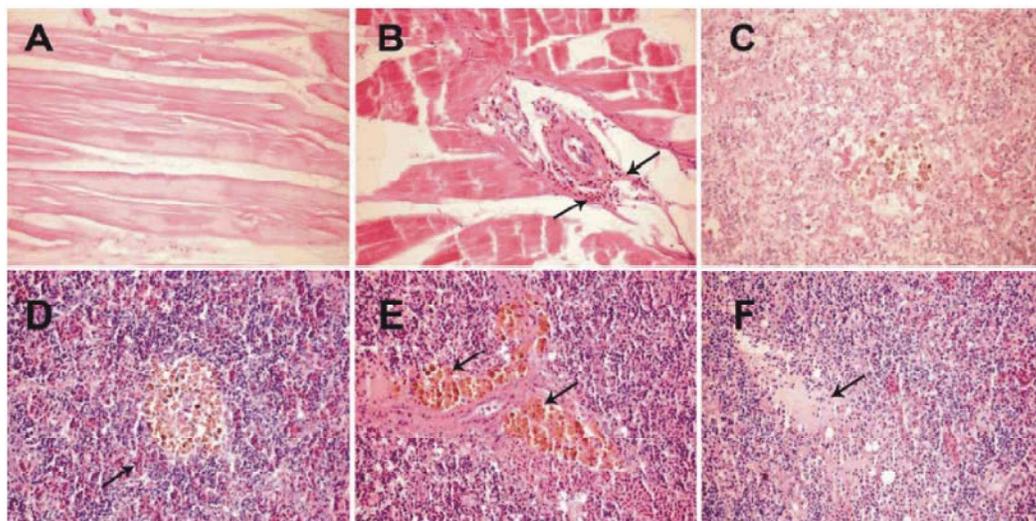


Fig. 6: Histological appearance of different tissues in *O. niloticus*. Photomicrographs of muscle tissues; (A) no histopathological changes (B) perivascular oedema and hemorrhage Photomicrographs of spleen tissues; (C) normal structure (D) marked congestion associated with diffuse infiltration with inflammatory cells (E) hyperactivity of melanomacrophage centers around blood vessels; (F) focal area of necrosis. All tissues stained with hematoxylin and eosin (x400). A and C photomicrographs represented reference area, while the others represented polluted area.

lamellar deformations with prominent abnormal arrangement, congestion and necrosis of gills' lamellae (Fig. 5B-C). Microscopical examination of kidney sections of tilapias collected from the reference area revealed the presence of normal renal corpuscles, glomerular tufts and renal tubules (Fig. 5D). In polluted area, histopathological examination of sampled tilapias' kidney tissues revealed an interstitial nephritis and hyalinosis of the renal blood vessels' wall. The tubular system exhibited focal necrosis which further replaced by massive inflammatory cells and free RBCs (Fig. 5 E,F,G) together with interstitial nephritis in some renal portions. Hepatic tissue of fishes collected from polluted area have presented vacuolar degeneration with hepato-pancreatic necrosis (Fig 5H) compared to Normal liver section reference area (Fig. 5I). Muscle tissues of tilapias collected from Area 2 showed perivascular edema and hemorrhagic myositis (Fig. 6B) compared to normal arrangement of muscle bundles in case of those fishes from reference area (Fig. 6A). Splenic tissues of reference area fishes have demonstrated normal organization of splenic compartments (Fig. 6C). However, the splenic tissues of Area 2 showed activation of melanomacrophage centers around blood vessels (Fig. 6D). Splenic tissues were congested, necrotic together with diffuse infiltration of inflammatory cells (Fig. 6E,F).

DISCUSSION

Alterations in immunological functions of an aquatic organism can be used as immunological indicators for evaluating the direct effects of exposure to any pollutant [24]. The external barriers separating the fish from its environment, i.e., the epithelia of skin, gills and alimentary canal, is the first line of defense. These epithelia work as mechanical barriers to invading pathogens, but they also contain chemical (antibodies, lysozyme, etc.) and cellular (immune cells) defenses. In fact, a wide variety of chemicals has been reported to impact immune parameters of teleost fishes [25,26].

According to Egyptian Chemical Standards, the maximum permissible limits of heavy metal in surface water are 5 mg for Zn, 1 mg/l for Cu, 0.01 mg/l for Cd, 0.05 mg/l for Pb and 1 mg/l for Fe [27]. Comparable U.S. Environmental Protection Agency (US-EPA) standards are 5 mg for Zn, 0.5 mg/l for Mn, 1 mg/l for Cu, 0.01 mg/l for Cd, 0.005 mg/l for Pb and 0.3 mg/l for Fe [28]. In the current study, the water samples from Area 2 were polluted with Pb and Fe which exceeded the permissible limits according to US-EPA and ESC. In contrast, levels of Zn, Cu, Mn and Cd from both sites are within the permissible limits according to US-EPA [28], ESC [27] and WHO [29] standards. This pattern of heavy metal traces

in water of predictably polluted area was consistent with the levels of heavy metals in water samples from the river Tributaries in Egypt [30]. Cadmium and Pb concentrations were comparable with Gomaa *et al.* [31] from Greater Cairo and with Monroy *et al.* [32] in Lake Titicaca. However, lower concentrations in raw water samples from Greater Cairo in Egypt were reported by Mohamed and Osman [33]. In polluted area, there are industrial activities that can introduce Cd and Pb to River Nile water. In addition, discharge of various treated and untreated liquid wastes to the water can introduce large quantities of heavy metals to the River Nile [34].

Interestingly, the fish collected from the industrial zone showed a significant increase in the total proteins. This result was in full agreement with previous study of Salah El-Deen *et al.* [35]. The increase in serum total protein level in fish may be attributed to several biological conditions as damage of liver, kidney and gills [36]. Also, elevation in serum protein level was reported by Ghazaly and Said [37] in case of *O. niloticus* exposed to sub-lethal concentration of Cu (1.14 mg/l) for 96 h. Herein, the increase of total proteins is associated with the reduction of total globulins. It is reported that Cd treatment inhibited the antibody levels in cunners (*Tautoglabrus adspersus*) [38].

Lysozyme is a major component of the fish innate immune system involved in inflammatory processes [39] acting against parasitic, bacterial and viral infections [40]. In the present study, the lysozyme activity was in agreement with Zelikoff *et al.* [41] who have concluded that lysozyme activity was not affected in serum of rainbow trout exposed to similar types of pollution. In other studies, lysozyme levels were reduced in the kidney of carp [42]. It is possible that heavy metals had a limited effect on lysozyme activity in the Nile Tilapia.

The principal cells involved in phagocytosis process are tissue macrophages, neutrophils and monocytes in blood [6]. The main functions of the phagocytic cells are to phagocytose tissue debris and microorganisms, to mount immune response regulating mediators and to bridge innate and adaptive immune responses. In the current work, there was a marked reduction in phagocytosis percentage in polluted area. Our results contradicts other studies documenting that zebra fish Cu exposure suppressed phagocytosis in European sea bass [43] and *Daniorerio* [44]. Similar results were also demonstrated in experimental fish [45].

It has been implied that there is an association between DNA damage in aquatic animals and contamination of aquatic environment [46,47]. In the

present assay, we found that the DNA strand breaks increased statistically in tilapia collected from polluted area compared to fish collected from reference area. Likewise, comet fish exposed subchronically and chronically to effluents from a Swine industry associated with greater DNA damage [48]. Increased number of DNA strand breaks correlated with immunotoxicity in dolphins [49]. Further, Erythrocytes genotoxicity of *O. niloticus* exposed to polluted river water samples was most recently confirmed in Fuzinato *et al.* [50] study.

In the literature, heavy metal concentrations in the tissue of freshwater fish vary considerably among different studies. The accumulation patterns of contaminants in fish depend on both uptake and elimination rates [51]. In the current study, the low accumulation of Cu in gills may be due to development of some defensive mechanism such as excessive mucous secretion and clogging of gills which runs parallel with similar explanation of Eissa *et al.* [4] who described similar mechanisms of gills of Nile tilapias and catfish that have survived a colossal environmental crisis of Mariotteya water body in 2009. Additionally, Cu can combine with other contaminants such as ammonia, mercury and Zn to produce an additive toxic effect on fish. The present results agreed with previous studies which found that Cu and Zn showed higher concentrations in liver than flesh muscles of different fish species from Egypt and different places of the world [31, 52]. Recently, it has been pointed to Cu rapid deposition in muscles at early exposure, but liver acted as a terminal Cu storage area at prolonged exposure [53]. Our data revealed that accumulation pattern of the trace heavy metals in fish liver from contaminated area were in comparable with liver of *O. niloticus* obtained from industrial area of Shoubra El-Khema [54].

Cd and Pb are toxic at low concentrations, non-biodegradable non-essential heavy metals and have no role in biological processes in living organisms. Thus, even in low concentration, they could be harmful to fish. The obtained data illustrated that pattern of Cd and Pb accumulation in different tissues were in agreement with Rashid [55] who indicated that the highest concentration of Pb in fish were kidney and liver followed by bone and muscles of *O. niloticus* from Nile River at Assiut region. Our results showed that the Mn is highly accumulated in gills than other organs of fish collected from both areas. These results are in accordance with Osman and Kloas [56] who demonstrated that Mn was highly concentrated in the gills of fish from early six sites along the whole River Nile. Fe is an abundant and important element, unsurpassed by any other heavy

metals in the earth's crust [54]. The increase of Fe accumulation in fish in all examined tissues was higher than the other metals possibly due to the increase of total dissolved Fe in Nile water and consequently increases the free metal Fe concentration and there by lead to an increase in metal uptake by different organs [57]. Further, the presence of the regional Iron and steel factory in El-Tebeen and its chronic dumping of factory effluents into River Nile area could be another staggering cause of such elevation of Fe levels in water as well as fish tissues. Interestingly, muscles were found to accumulate the lowest levels of heavy metal residues for all investigated metals in fish muscles from the studied both sites. Such tiny concentrations of heavy metals might have reached tissues through circulation [56].

Histological examination of liver showed increased vacuolation with congestion of blood vessels and necrosis of hepatocytes which may be attributed to the cumulative effect of heavy metals and the increase of their concentrations in the hepatic tissue. These results agreed with several previous studies [58]. In addition, the pathological liver steatosis was reported by Kranz and Peters [59], accompanied by additional features such as necrosis, lymphocyte infiltration, deposition of ceroids and aggregations of macrophages. The current histological study of the gills of *O. niloticus* in polluted area showed several histological alterations, including lamellar deformations, congestion and necrosis compared to reference area. It is possible that the damage of the gills could be a direct result of the heavy metals [60]. It is previously indicated to the presence of infiltrations of inflammatory cells in secondary cells [61]. Mucinous metaplasia of lamellar epithelial lining is considered as adaptive mechanism against heavy metal toxicity.

CONCLUSION

The results presented in this work suggested the possible immunotoxic and genotoxic effects of heavy metals on Nile tilapia. These data implied into the importance of immunological assays as environmental monitoring especially in industrial effluents.

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REFERENCES

1. Korium, M.A. and M.E.F. Toufeek, 2008. Studies of some physicochemical characteristics of old Aswan Dam Reservoir and River Nile water at Aswan. *Egyptian Journal of Aquatic Research*, 34: 149-167.
2. Gunkel, G., J. Kosmol, M. Sobral, H. Rohn, S. Montenegro and J. Aureliano, 2007. Sugar cane industry as a source of water pollution - case study on the situation in Ipojuca River, Pernambuco, Brazil. *Water, Air and Soil Pollution*, 180: 261-269.
3. Eissa, A.E., M. Moustafa, I.N. El-Husseiny, S. Saeid, O. S aleh and T. Borhan, 2009. Identification of some skeletal deformities in some freshwater teleost raised Egyptian aquaculture. *Chemosphere*, 77: 419-425.
4. Eissa, A.E., N.A. Tharwat and M.M. Zaki, 2013. Field assessment of the mid winter mass kills of trophic fishes at Mariotteya stream, Egypt: chemical and biological pollution synergistic model. *Chemosphere*, 90: 1061-1068.
5. Uribe, C., H. Folch, R. Enriquez and G. Moran, 2011. Innate and adaptive immunity in teleost fish: a review. *Veterinary Medicine*, 56: 486-503.
6. Magnadottir, B., 2006. Innate immunity of fish (overview). *Fish and Shellfish Immunology*, 9: 291-308.
7. Miller, G.G., L.I. Sweet, J.V. Adams, G.M. Omann, D.R. Passino-Reader and P.G. Meier, 2002. *In vitro* toxicity and interactions of environmental contaminants (Arochlor 1254 and mercury) and immunomodulatory agents (lipopolysaccharide and cortisol) on thymocytes from lake trout (*Salvelinus namaycush*). *Fish and Shellfish Immunology*, 13: 11-26.
8. Girón-Pérez, M.I., A. Santerre, F. Gonzalez-Jaime, J. Casas-Solis, M. Hernández-Coronado, Peregrina-Sandoval, J., A. Takemura and G. Zaitseva, 2007. Immunotoxicity and hepatic function evaluation in Nile tilapia (*Oreochromis niloticus*) exposed to diazinon. *Fish and Shellfish Immunology*, 23: 760-769.
9. Eissa, A.E., H.A. Hussein and M.M. Zaki, 2012. Detection of Avian Influenza (H5N1) insome fish and shellfish from different aquatic habitats across some Egyptian provinces. *Life Science Journal*, 9: 2702-2712.
10. Koca, S., Y.B. Koca, S. Yildiz and B. Gürcü, 2008. Genotoxic and histopathological effects of water pollution on two fish species, *Barbus capito pectoralis* and *Chondrostoma nasus* in the Büyük Menderes River, Turkey. *Trace Element Research*, 122: 276-291.

11. Jiang, D., Z. Hu, F. Liu, R. Zhang, B. Duo, J. Fu, Y. Cui and M. Li, 2014. Heavy metals levels in fish from aquaculture farms and risk assessment in Lhasa, Tibetan Autonomous Region of China. *Ecotoxicology*, 23: 577-583.
12. Karadede-Akin, H. and E. Unlu, 2007. Heavy metals concentrations in water, sediment, Fish and some benthic organisms from Trigris River, Turkey. *Environmental and Monitoring Assessment*, 131: 323-337.
13. Kaoud, H.A. and A.R. El-Dahshan, 2010. Bioaccumulation and histopathological alterations of the heavy metals in *Oreochromis niloticus* fish. *Natural Sciences*, 8: 147-156.
14. Tort, L., J.C. Balasch and S. Mackenzie, 2003. Fish immune system. A crossroads between innate and adaptive responses. *Immunology*, 22: 277-286
15. Stoskopf, M., 1993. *Fish Medicine*. WB Saunders Company, Philadelphia, Pennsylvania.
16. APHA (American Public Health Association), 2005. *American Water Works Association: Standard methods for the examination of water and wastewater*. New York.
17. Hseu, Z.Y., 2004. Evaluating heavy metal contents in nine composts using four digestion methods. *Bioresource Technology*, 95: 53-59.
18. Burtis, C.A., 1994. Methods for the determination of proteins in serum and plasma. In: *Textbook of Clinical Chemistry*, Eds., Burtis, C.A., R. Edward, M.D. Ashwoodand N.W. Tietz, W.B. Saunders Company, USA, 2ndedn., pp: 692-698.
19. Rouse, B.T., L.A. Babiuk and P.M. Henson, 1980. Neutrophils in antiviral immunity: inhibition of virus replication by mediators produced by bovine neutrophils. *The Journal of Infectious Diseases*, 141: 223-232.
20. Woldehiwet, Z. and T.G. Rowan, 1990. Some observation on the effects of age of calves on the phagocytosis and killing of *Staph. aureus* by polymorphonuclear leucocytes. *British Veterinary Journal*, 146: 165-170.
21. Schultz, L.A., 1987. *Methods in clinical chemistry*. The C.V. Mosby Co. St. Louis, pp: 742-746
22. Tice, R.R.1., E. Agurell, D. Anderson, B. Burlinson, A. Hartmann, H. Kobayashi, Y. Miyamae, E. Rojas, J.C. Ryuand and Y.F. Sasaki, 2000. Single cell gel/comet assay: guidelines for in vitro and in vivo genetic toxicology testing. *Environmental and Molecular Mutagenesis*, 35: 206-221.
23. Prophet, E., B. Mills and J. Arrington, 1992. *Laboratory Methods in Histotechnology*, 1st edn. American registry of Pathology, Washington DC.
24. Reynaud, S. and P. Deschaux, 2006. The effects of polycyclic aromatic hydrocarbons on the immune system of fish: a review. *Aquatic Toxicology*, 77: 229-238.
25. Bols, N.C., J.L. Brubacher, R.C. Ganassin and L.E. Lee, 2001. Ecotoxicology and innate immunity in fish. *Developmental and Comparative Immunology*, 25: 853-873.
26. Carlson, E. and J. Zelikoff, 2008. The immune system of fish: a target organ for toxicity. In *The toxicology of Fishes*, CRC Pres, Boca Raton, FL., Eds, Di Giulio, R., D.E. Hinton, pp: 489-530.
27. ECS (Egyptian Chemical Standards), 1994. *Protection of the Nile River and Water Stream from pollution*, Ministry of Irrigation, Cairo, Egypt, Law No 4.
28. US-EPA (United States Environmental Protection Agency), 2000. *Guidance for assessing chemical contaminant data for use in fish advisories*, Vol. 1. *Fish sampling and analysis* 3rdedn, office of science and Technology office of water USPA Washington, DC EPA 823-00-007, pp: 1-200.
29. WHO (World Health Organization), 1993. *Guidelines for drinking-water quality*, 2ndedn, Vol 1. *Recommendations*, Geneva.
30. Medhat, F.M. and I. Nasr, 2012. Metals in water from the Nile River Tributaries in Egypt. *Bulletin of Environmental Contamination and Toxicology*, 88: 594-596.
31. Gomaa, M.N.E., A.A.K. Abou-Arab, A. Badawy and K. Nauguib, 1995. Distribution pattern of some heavy metals in Egyptian fish organs. *Food Chemistry*, 53: 385-389.
32. Monroy, M., A. Maceda-Veiga and A. De Sostoa, 2014. Metal concentration in water, sediment and four fish species from Lake Titicaca reveals a large-scale environmental concern. *Science of the Total Environment*, 487: 233-244.
33. Mohamed, M.A.M. and M.A. Osman, 1998. Lead and cadmium in Nile River water and finishing drinking water in Greater Cairo, Egypt. *Environment International*, 24: 767-772.
34. Palanques, A., 1994. Distribution and heavy metal pollution of suspended particular matter on the Barcelont continental shelf (North-western Mediterranean. *Environmental Pollution*, 85: 205-215.

35. Salah El-Deen, M.A., K.M. Sharada and S.M. Abdu El-Ella, 1996. Some metabolic alternation in grass carp (*Ctenopharyngodon idella*) induced by exposure to cadmium. Journal of the Egyptian German Society of Zoology, 21: 441-457.
36. Gluth, G. and W. Hanke, 1985. A comparison of physiological changes in carp, *Cyprinus carpio* induced by several pollutants of sublethal concentrations. I- The dependency on exposure time. Ecotoxicology Environmental Safety, 9: 179-188.
37. Ghazaly, K.S., 1992. A comparative study of trace elements accumulation in tissues of the teleost *Tilapia zillii* from contaminated and clean areas. Bulletin of the National Institute of Oceanography and Fisheries, 18: 37-41.
38. Robohm, R.A., 1986. Paradoxical effects of cadmium exposure on antibacterial antibody responses in two fish species: inhibition in cunners (*Tautoglabrus adspersus*) and enhancement in striped bass (*Morone saxatilis*). Veterinary Immunology and Immunopathology, 12: 251-262.
39. Bayne, C.J. and L. Gerwick, 2001. The acute phase response and innate immunity of fish. Developmental and Comparative Immunology, 25: 725-743.
40. Fatima, M., S.N. Mandiki, J. Douxfils, F. Silvestre, P. Coppe and P.Kestemont, 2007. Combined effects of herbicides on biomarkers reflecting immune-endocrine interactions in goldfish. Immune and antioxidant effects. Aquatic Toxicology, 81: 159-167.
41. Zelikoff, J.T., D. Bowser, K.S. Squibb and K. Frenkel, 1995. Immunotoxicity of low level cadmium exposure in fish: an alternative animal model for immunotoxicological studies. Journal of Toxicology and Environmental Health, 45: 235-248.
42. Sövényi, J. and J. Szakolczai, 1993. Studies on the toxic and immunosuppressive effects of cadmium on the common carp. Acta Veterinaria Hungarica, 41: 415-426.
43. Bennani, N., A. Schmid-Alliana and M. Lafaurie, 1996. Immunotoxic effects of copper and cadmium in the sea bass *Dicentrarchus labrax*. Immunopharmacology and Immunotoxicology, 18: 129-144.
44. Rougier, F., D. Troutaud, A. Ndoye and P. Deschaux, 1994. Non-specific immune response of Zebrafish, *Brachydanio rerio* (Hamilton-Buchanan) following copper and zinc exposure. Fish and Shellfish Immunology, 4: 115-127.
45. Lugo, R.S., G. Nathali, L.B. Villalobos de B and L. Mairin, 2006. Immunological response of the freshwater fish *Colossomama cropomum* as a biomarker of copper exposure. Bulletin of Environmental Contamination and Toxicology, 77: 925-930.
46. Klobucar, G.I., A. Stambuk, M. Pavlica, M. Sertić Perić, B. Kutuzović Hackenberger and K. Hylland, 2010. Genotoxicity monitoring of freshwater environments using caged carp (*Cyprinus carpio*). Ecotoxicology, 19: 77-84.
47. Fatima, M., N. Usmani, M. Mobarak Hossain, M.F. Siddiqui, M.F. Zafeer, F. Firdaus and S. Ahmad, 2014. Assessment of Genotoxic Induction and Deterioration of Fish Quality in Commercial Species Due to Heavy-Metal Exposure in an Urban Reservoir. Archives of Environmental Contamination Toxicology, 67: 203-213.
48. Lima, P.L., J.C. Benassi, R.C. Pedrosa, J. Dal Magro, T.B. Oliveira and D. Wilhelm Filho, 2006. Time-course variations of DNA damage and biomarkers of oxidative stress in tilapia (*Oreochromis niloticus*) exposed to effluents from a swine industry. Archives of Environmental Contamination Toxicology, 50: 23-30.
49. Lee, R.F. and S. Steinert, 2003. Use of the single cell gel electrophoresis/comet assay for detecting DNA damage in aquatic (marine and freshwater) animals. Mutation Research, 544: 43-64.
50. Fuzinato, C.F., L. Flohr, S.P. Melegari and W.G. Matias, 2013. Induction of micronucleus of *Oreochromis niloticus* exposed to waters from the Cubatão do Sul River, southern Brazil. Ecotoxicology and Environmental Safety, 98: 103-109.
51. Hakanson, L., 1984. Metals in fish and sediment from the River Kolbacksan water system, Sweden. Archives of Hydrobiology, 101: 373-400.
52. Eisenberg, M. and J.J. Topping, 1986. Trace metal residues in finfish from Maryland waters, 1978- 1979. Journal of Environmental Science and Health, 21: 87-102.
53. Tsai, J.W., Y.R. Ju, Y.H. Huang, Y.S. Deng, W.Y. Chen, C.C. Wu and C.M. Liao, 2013. Toxicokinetics of tilapia following high exposure to waterborne and dietary copper and implications for coping mechanisms. Environmental Science and Pollution Research International, 20: 3771-3780.

54. El-Naggar, A.M., S.A.Mahmoud and S.I. Tayel, 2009. Bioaccumulation of some heavy metals and histopathological alterations in liver of *Oreochromis niloticus* in relation to water quality at different localities along the River Nile, Egypt. World Journal of Fish and Marine Sciences, 1: 105-114.
55. Rashid, M.N., 2001. Cadmium and lead levels in fish (*Tilapia nilotica*) tissues as biological indicator for Lake water pollution. Environ. Environmental and Monitoring Assessment, 68: 75-89.
56. Osman, A.G.M. and W. Kloas, 2010. Water quality and heavy metal monitoring in water, sediments and tissues of the African Catfish *Clarias gariepinus* (Burchell, 1822) from the River Nile, Egypt. Journal of Environmental Protection, 1: 389-400.
57. Tayel, S., A.M. Yacoub and S. Mahmoud, 2008. Histopathological and haematological responses to freshwater pollution in the Nile Catfish *Clarias gariepinus*. Journal of Egyptian Academic Society and Environmental Development, 9: 43-60.
58. Abd El-Gawad, A.M., 1999. Histopathological studies on the liver and gills of Nile tilapia, (*Oreochromis niloticus*) exposed to different concentrations of lead acetate and zinc sulphate. Journal of the Egyptian German Society of Zoology, 30: 3-22.
59. Kranz, H. and N. Peters, 1985. Pathological conditions in the liver of ruffe *Gymnocephalus cernua* (L.), from the Elbe estuary. Journal of Fish Diseases, 8: 13-24.
60. El-Bakary, N.E.R., S.B. Said and A. El-Badaly, 2011. Using *Oreochromis niloticus* for assessment of water quality in water treatment plants. World Applied Sciences Journal, 12: 1455-1463.
61. Randi, A.S., J.M. Monserrat, E.M. Rodrigue and L.A. Romano, 1996. Histopathological effects of cadmium on the gills of the freshwater fish, *Macropsobrycon uruguayanae Eigenmann* (Pisces, Atherinidae). Journal of Fish Diseases, 19: 311-322.